

Challenges for the sustainable use of water and land resources under a changing climate and increasing salinization in the Jizzakh irrigation zone of Uzbekistan (Postprint)

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Abstract

Jizzakh Province in Uzbekistan is one of the largest irrigated areas in Central Asia without natural drainage. In combination with aridity, climate change and extensive irrigation practices, this has led to the widespread salinization of agricultural land. The aim of this study was to identify opportunities to improve the reclamation status of the irrigated area and how best to effectively use the water resources in Jizzakh Province based on investigations conducted between 1995 and 2016. A database of field measurements of groundwater levels, mineralization and soil salinity conducted by the provincial Hydro-Geological Reclamation Expeditions was used in the study. The total groundwater mineralization was determined using a portable electric conductometer (Progress 1T) and the chloride concentration was determined using the Mohr method. The soil salinity analyses were conducted by applying two different methods: (1) the extraction and assessment of the soluble salt content, and (2) using an SM-138 conductivity sensor applied to a 1:1 mixture of soil sample and water. The analyses of the monitoring results and the salt balance in the “irrigation water-soil-drainage water” system clearly demonstrated that the condition of the irrigated land in the province was not significantly improved. Under these conditions, the stability of crop yields is achieved mainly through the use of large volumes of fertilizer. However, excess amounts of mineral fertilizers can also cause the salinization of soils. The average groundwater salinization value in most of the irrigated land (75.3%) fluctuated between 1.1 and 5.0 g/L, while the values were less than 1.0 g/L in 13.1% of the land and in the range of 5.1-10.0 g/L in 10.5% of the land. During the period of 1995-2016 the salinization level of the irrigated land in Jizzakh Province increased slightly and the area could be divided into the following classes: no salinity (17.7% of the total area), low salinity (51.3%),

moderate salinity (29.0%), and high salinity (2.0%). Detailed studies of the salt balance in irrigated land, the impact of climate change, increased fertilizer use, and repeated remediation leaching on the groundwater level and mineralization should be conducted in the future, due to the possibility of accelerated salinization, fertility decline, and reduced yields of agricultural crops.

Full Text

Preamble

Challenges for the Sustainable Use of Water and Land Resources under a Changing Climate and Increasing Salinization in the Jizzakh Irrigation Zone of Uzbekistan

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Abstract: Jizzakh Province in Uzbekistan is one of the largest irrigated areas in Central Asia without natural drainage. In combination with aridity, climate change, and extensive irrigation practices, this has led to widespread salinization of agricultural land. The aim of this study was to identify opportunities to improve the reclamation status of the irrigated area and determine how to effectively use water resources in Jizzakh Province based on investigations conducted between 1995 and 2016. A database of field measurements of groundwater levels, mineralization, and soil salinity conducted by the provincial Hydro-Geological Reclamation Expeditions was used in the study. Total groundwater mineralization was determined using a portable electric conductometer (Progress 1T) and chloride concentration was determined using the Mohr method. Soil salinity analyses were conducted using two different methods: (1) extraction and assessment of soluble salt content, and (2) an SM-138 conductivity sensor applied to a 1:1 mixture of soil sample and water. Analyses of monitoring results and the salt balance in the “irrigation water–soil–drainage water” system clearly demonstrated that the condition of irrigated land in the province was not significantly improved. Under these conditions, the stability of crop yields is achieved mainly through the use of large volumes of fertilizer. However, excess amounts of mineral fertilizers can also cause soil salinization. The average groundwater salinization value in most of the irrigated land (75.3%) fluctuated between 1.1 and 5.0 g/L, while values were less than 1.0 g/L in 13.1% of the land and in the range of 5.1–10.0 g/L in 10.5% of the land. During the period 1995–2016, the salinization level of irrigated land in Jizzakh Province increased slightly, and the area could be divided into the following classes: no salinity (17.7% of the

total area), low salinity (51.3%), moderate salinity (29.0%), and high salinity (2.0%). Detailed studies of the salt balance in irrigated land, the impact of climate change, increased fertilizer use, and repeated remediation leaching on groundwater level and mineralization should be conducted in the future due to the possibility of accelerated salinization, fertility decline, and reduced yields of agricultural crops.

Keywords: irrigation; groundwater level; salinity; soil salinization; salt balance; Uzbekistan

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1 Introduction

Agriculture is an important economic sector in Uzbekistan, accounting for about 23% of GDP and 27% of total employment (Kulmatov, 2014). Irrigated farming produces more than 90% of agricultural yield and consumes more than 90% of the nation's available water resources annually (Uzbekistan N C, 2009; Dyhovniy and Schutter, 2011; Opp et al., 2016). However, due to the flat topography of many irrigated areas in the Aral Sea Basin (ASB) countries, particularly the Republic of Uzbekistan, which has a low level of natural drainage and extensive areas of flood irrigation with poor-quality water, soil salinization and degradation of irrigated land has occurred (Panin, 1958; Pankova et al., 1996; Gafurova et al., 2005; SCNP, 2008; Kulmatov, 2018).

Currently, more than 60% of all irrigated areas in Uzbekistan are affected by salinization, which has resulted in reduced crop yields (e.g., 20%-30% for cotton). In most farms, increased reclamation of leached water and application of large quantities of mineral fertilizers usually overcome these losses (SCNP, 2008; Kulmatov, 2014, 2018). Agricultural, industrial, and urban development have all contributed to increased salinity in streams and rivers, but the likely effects of future development and climate change are unknown. Most of the change is related to the extent of human land uses, with climate change accounting for only 12% of the increase (Olson, 2019). However, land salinization and degradation are aggravated by the vulnerability of sensitive arid lowlands to climate change.

Uzbekistan is located in an arid plain and the semi-arid foothill areas of the endorheic (closed) ASB. Being the main water consumer in the ASB, it receives about 80%-85% of its water resources from neighboring upstream countries, primarily through three transboundary rivers: the Syrdarya, the Amudarya, and the Zarafshan (FAO, 2017). For these and many other rivers of the ASB, wa-

ter from snowmelt and glaciers in the Pamir and Tian Shan mountains is the main source of river discharge, and therefore their runoff is strongly dependent on changes in precipitation patterns and seasonal snow water storage capacity. The amount of precipitation received in recent years has decreased across the entire region, especially in the western parts of Turkmenistan, Uzbekistan, and Kazakhstan (Bai et al., 2011; Lioubimtseva et al., 2018). Endorheic basins are known to be sensitive to various changes in ambient conditions, which makes them vulnerable to climate change and other human-induced pressures, including agricultural intensification (Karthé et al., 2017). Water abstraction for irrigation is the main cause of reduced river flows in downstream areas, but climate change also leads to decreased water discharges in river catchments. Temperature increases followed by a spread in aridity have been observed during the last century throughout Central Asia, and this is predicted to continue in the future at a rate above the global average (Lioubimtseva and Henebry, 2009; Mannig et al., 2013). There is already clear evidence of negative consequences of climate change, including decreased available water resources and declining agricultural productivity. Between the 1960s and the 2010s, the glacierized area decreased by 23%-49% in various river catchment basins of the ASB (Semakova et al., 2015). During the period 2001-2010, the estimated volume of total glacier runoff in the Amudarya and Syrdarya river basins was reduced from 19.0 and 3.4 km³ to 18.0 and 3.2 km³, respectively (Savoskul and Smakhtin, 2013).

In the future, upstream countries are planning to construct facilities to accumulate additional water volumes in their existing water reservoirs for hydropower energy production, which may present another potential threat for land use sustainability. In combination with the high population growth rate in the ASB, this is likely to lead to decreased availability of river water and increased water consumption (Uzbekistan N C, 2009; Dukhovny and Schutter, 2011; Groll et al., 2015). These trends are likely to be exacerbated until coordinated mitigation and adaptation activities are implemented (Dukhovny and Schutter, 2011; Groll et al., 2015; Canedo-Arguelles et al., 2016).

Furthermore, the extensive area of unsustainable land use and insufficient maintenance of irrigation and drainage networks has caused an increase in groundwater level and mineralization. As a result, there has been a dramatic deterioration in the condition of irrigated areas in recent decades (SCNP, 2013; Kulmatov, 2014, 2018).

A comparative analysis of climatic characteristics of desert and arid lands in Central Asia (Turan Depression, the Gobi Desert, and the deserts of the Dzungar and Tarim depressions) revealed that the degree of aridity, the degree of continentality, and the precipitation amount and regime differ among these areas (Pankova and Konyushkova, 2013). The deficit of moisture in the modern climate results in the preservation of salt accumulations at their point of origin. The importance of specific features of the climate in these regions has been noted, including the precipitation regime and the effect of redistribution of salts in the profiles of automorphic salt-affected soils (Pankova and Konyushkova,

2013).

In recent years, there has been a deterioration in the quality and decrease in the quantity of surface water resources and reclamation conditions in the irrigated lands of provinces located in the mid- and downstream areas of the Amudarya River (Ibrakhimov et al., 2007). Only in Navoi Province in Uzbekistan between 2000 and 2015 has the groundwater level and soil salinity shown slight improvement (Kulmatov et al., 2018; Kulmatov et al., 2015).

In irrigated land in Jizzakh Province (Syrdarya River Basin), the amount of humus and the absorption capacity has declined in recent years, and simultaneously the mechanical composition and fertility of soils has decreased (Sherimbetov, 2015; Rakhmatov and Abdullaev, 2016). Some studies have shown that long-term irrigation has a significant impact on the absorption capacity and physical and chemical properties of soils. When agricultural activities are conducted in irrigated soils for long periods, there are varying degrees of change in the soil layers.

It has been revealed that of the four soil components (humus, carbonates, gypsum, and water-soluble salts), water-soluble salts play the dominant role in spectral reflectance from saline soils (Karavanova et al., 2001; Sherimbetov, 2015).

Previous studies have attempted to investigate the relationships between soil productivity and salinization to ensure sustainable use of water and land resources in the study area. Unfortunately, most of these studies were conducted within a short period of time, and the ionic composition and mineralization of irrigation water over long periods were not sufficiently taken into account. The salt balance in the “irrigation water–soil–drainage water” system has not been studied. The pathways of salt migration in natural and agricultural land are important to understand for long-term monitoring and prevention of soil and water salinization; however, they have still not been quantitatively determined in Jizzakh Province.

The objective of this study was to determine the long-term changes (1995–2016) in the quality of water used for irrigation, the level and mineralization of groundwater, the amelioration required in irrigated land, and the salt balance in the “irrigation water–soil–drainage water” system, as well as to develop recommendations for sustainable use of water and land resources, with a specific focus on the Jizzakh irrigation zone.

2 Study Area

Jizzakh Province is located in the center of the Republic of Uzbekistan. The province is bordered by Syrdarya Province in the east, the Republic of Kazakhstan in the north, and the Republic of Tajikistan in the south, and is enclosed by the Syrdarya River, the Turkestan border, and the Kizil-Kum Desert. The

total area of irrigated land (39°57' -41°24' N, 66°66' -68°57' E) in the province is 3.0×10^6 hm², and the main agricultural crops are cotton, wheat, berries, fruit, grapes, melons, and gourds.

The province is situated in a continental area characterized by dry and hot summers and moderately cold winters. Precipitation occurs mainly in winter and spring, with an average annual rainfall of 308 mm. The highest recorded air temperature was 36.4°C, observed in July. The average annual air temperature is 15.6°C. The humidity level is 70%–80% in winter. The vegetation growth period is 210–240 days (State Department of Statistics of Uzbekistan, 2017).

The soils of Jizzakh Province are classified as light gray meadow-brown, dark-brown, typical and light-brown grass, meadow-grass, grassland and swampy-grass, sandy desert, sand, and sandy-saline soil (SCNP, 2016). The soils in the irrigation zone are classified as typical dark-brown and light-brown, meadow-grassy, grass-grassland, and meadow and wetland-grass soils.

The water supply in Jizzakh Province is transboundary and largely dependent on the quantity of water coming from countries containing the upper reaches of the main river systems, i.e., Kyrgyzstan and Tajikistan. There are three main sources of irrigation water in the province with average annual volumes of diverted water of 1.89 km³ (62.2% of the total) from the Syrdarya River, 1.05 km³ (34.5%) from the Zarafshan River, and 0.10 km³ (3.3%) from the Sangzor River. The total water volume used in the province is 3.04 km³, of which 2.85 km³ (93%) is used for agricultural purposes and the remaining 0.19 km³ (7%) is used by industry and for drinking water supply.

The organic matter concentration in water from the Syrdarya and Zarafshan rivers is very low; however, both rivers have elevated concentrations of inorganic substances, mainly in the form of sulfate, chloride, and carbonate ions (Table 1), which causes secondary salinization of agricultural land (Yakubov et al., 2011; Karimov et al., 2014; Karimov et al., 2019).

Table 1 Mineralization of river water used for irrigation in Jizzakh Province during 2000–2016

River	Hydropost	River water mineralization (g/L)
		Min
Syrdarya	Nadejdinskiy	
Zarafshan	Pervomayskiy	

Note: Min, minimum; Max, maximum; Ave, average. Hydropost -measuring infrastructure within the main canal.

3 Materials and Methods

The study used a database of field measurements of groundwater, mineralization, and soil salinity levels conducted by the provincial Hydro-Geological Reclamation Expeditions (HGRE) under the Uzbek Basin Irrigation System Administration of the Ministry of Agriculture and Water Resources of the Republic of Uzbekistan. The HGRE conducts a monitoring program throughout Uzbekistan.

Indicators of spatial and temporal changes in the levels and mineralization of groundwater in irrigated areas of Jizzakh Province were determined by HGRE specialists through the use of 2,255 continuous monitoring wells. The Ministry of Agriculture and Water Resources also has a detailed database of long-term inorganic fertilizer use in all irrigated land in Uzbekistan, including Jizzakh Province, which has been processed statistically and can be used to study the dynamics of fertilizer use. This existing extensive database was complemented by additional data regarding soil characteristics, soil salinity, and cropping patterns based on the results of a soil-sampling exercise conducted by the authors.

Groundwater samples were taken by on-farm technicians, allowing the swift collection and analysis of more than 12,600 groundwater samples per year. The groundwater level monitoring stations were equipped with a tube (inner diameter of 90–110 mm, length of 3–6 m) filled with sand-gravel filters. Samples of groundwater from 0.0–1.0, 1.0–1.5, 1.5–2.0, 2.0–3.0, 3.0–5.0, and 10.0 m deep monitoring wells were taken and sent to the laboratory for analysis.

The determination of groundwater level and mineralization was performed three times a year on April 1, July 1, and October 1. The total mineralization of groundwater was determined using a portable electric conductometer (Progress 1T, Central Asian Scientific Research Institute of Irrigation (SANIIRI), Uzbekistan) and the chloride concentration was determined using the Mohr method, i.e., argentometric titration. During the first sampling period (April), the effects of extensive salt leaching, which occurred just before the irrigation season started, could be detected. The second sampling period (July) covered the period of peak irrigation activity, and the third sampling period (October) took place immediately after the end of the growing season. This enabled analysis of the lowering of the phreatic surface without groundwater recharge. An analysis of groundwater level dynamics outside the growing season is important because seasonal salinity restoration might occur when the upward flux prevails over the lateral outflow.

During 2000–2016, 111,634 groundwater samples were analyzed. Groundwater mineralization was assessed and classified following the approach of Priklonsky (1970) (Table 2).

Table 2 Classification of groundwater (GW) based on total mineralization (Priklonsky, 1970)

Category	Total dissolved solids (TDS; g/L)
Fresh	
Low mineralization	
Medium mineralization	
High mineralization	

A total of 17,500 soil samples was taken each year by HGRE staff at the end of the growing season in November during 2000–2016. Each soil sample consisted of three subsamples collected at depths of 0–30, 30–70, and 70–100 cm. More than 116,000 soil samples were collected from these depths in the irrigated areas to determine soil salinity. Each sample was considered representative of an area of 10–20 hm², resulting in a much denser grid of soil data than groundwater data.

Mineral fertilizers (mainly nitrogen-containing) can cause salinization of irrigated soils. In the calculation of the salt balance, the contribution of mineral fertilizers was included in the total balance of salts because their residues enter the collector-drainage network, and therefore their contribution was not taken into account separately. The effects of side flows, for example, unsaturated groundwater from irrigated land, were not taken into account when calculating the salt balance.

The soil salinity level of irrigated land was measured twice a year on April 1 and October 1. For each soil sample, the electric conductivity was measured at the four corners of a 1.5×2.0 m² area surrounding the soil-sampling site. The soil samples were dried naturally in a location with no direct sunlight. The soil samples were then passed through a special sieve with a mesh size of 0.1 cm, and three solution samples were prepared with 50 g of soil in distilled water and analyzed. The analyses of soil salinity were conducted by applying two different methods: first through the extraction and assessment of the soluble salt content and then through a conductometer (IKS Express 1T, SANIIRI, Uzbekistan) applied to a 1:1 mixture of soil sample and water (Shirokova and Chernyshev, 1999).

The measured soil salinity (total mineralization) was then categorized as low (0.3–1.0 g/L), moderate (1.0–2.0 g/L), high (2.0–3.0 g/L), or very high (>3.0 g/L). This classification was assigned to the area represented by each soil sample to obtain spatial and temporal information about soil salinity distribution and dynamics. Based on the groundwater and soil data, the overall extent of salinization of the irrigated areas was assessed using the salinity classification system developed by Bazilevich and Pankova (1970) (Table 3).

Table 3 Classification of soil salinity, based on TDS and chloride (Cl) concentration (modified after Bazilevich and Pankova (1970))

Level of salinization	TDS (g/L)	Cl (g/L)
	Sulfate	Chloride-sulfate
No salinity	<0.01	>0.30
Low salinity	<0.01	<0.01
Moderate salinity		
High salinity		
Very high salinity		

4.1 Groundwater Level and Its Influence on the Characteristics of Irrigated Land

In arid areas, factors affecting the efficiency and sustainability of irrigated agriculture are closely linked with groundwater level. Improper irrigation and inefficient drainage networks can lead to soil salinization (Eshchanov, 2008; Martin and Sauerborn, 2013; Kulmatov, 2018). To ensure sustainable long-term irrigation and crop yields, it is necessary to determine the reasons for temporal and spatial changes in GW level and mineralization. High groundwater salinity in combination with shallow groundwater level leads to salinization of irrigated land and to bogging of crop root systems, resulting in yield losses (Gafurova et al., 2005; Eshchanov, 2008).

Table 4 shows the admissible levels of groundwater salinity for different groundwater depths. A groundwater level of less than 1.0 m is especially critical because even salt concentrations of less than 1.0 g/L can lead to land salinization, while groundwater at depths of 3.0 m or more could have salinity of up to 5.0 g/L before the full salinization effect sets in (Gafurova et al., 2005; Kulmatov et al., 2018).

These “critical depths” were used to assess the level of amelioration required for irrigated land and to conduct practical measures against salinization of irrigated land. If groundwater was located above “the critical depth,” the upward movement of salts from the lower layers would increase and soils would be exposed to salinization. However, if groundwater was located below the critical depth, salt would not rise from the lower layers into the topsoil and soil salinization would not occur.

Table 4 The critical groundwater (GW) depth and the related admissible level of GW salinity in irrigated soils (Gafurova et al., 2005)

Critical GW depth (m)	Admissible level of GW salinity (g/L)	Maximum concentration of chlorine in GW (%)
About 1.0	>0.69	

The critical groundwater depth also depends on soil properties, especially the capillary structure, water retention potential, and percolation characteristics (e.g., the amount of macropores or soil density) (Childs, 1969; Nielsen et al., 1973; Bowles, 1979; Heath, 2004; Gafurova et al., 2005; Kulmatov, 2018). To take these parameters into account, Table 5 shows the critical groundwater depth in relation to the water retention and percolation capability of irrigated land. This integration of several important soil hydrological parameters makes the critical groundwater depth a crucial concept for sustainable use of irrigated land resources.

Table 5 Critical GW depth related to the water retention and percolation capability of irrigated soils (Gafurova et al., 2005)

Water retention and percolation capability of irrigated soils	GW salinity (g/L)	Critical GW depth (m)	Admissible GW level (m)
Average	Strong	Strong	

To ensure long-term sustainable irrigation in farming and stable crop yields, it is necessary to determine the reasons for temporary changes in groundwater level, with special consideration given to the critical groundwater level and salinity. An analysis of groundwater level dynamics in Jizzakh Province (Fig. 1 [Figure 1: see original paper]) revealed that there were no irrigated fields with very shallow (0.0–1.0 m) groundwater levels. The largest area of all irrigated land in the province ($200.0 \times 10^3 \text{ hm}^2$; 66.96% of total irrigated land) had a groundwater table at a depth of 2.1–3.0 m; followed by an area of $63.93 \times 10^3 \text{ hm}^2$ (21.41%) at 3.1–5.0 m; $2.71 \times 10^3 \text{ hm}^2$ (9.27%) at 5.0 m; $6.23 \times 10^3 \text{ hm}^2$ (2.09%) at 1.5–2.0 m; and $0.80 \times 10^3 \text{ hm}^2$ (0.27%) at 1.0–1.5 m. During 1995–2016, the groundwater level in a large area of the province's irrigated land ($291.00 \times 10^3 \text{ hm}^2$) remained generally stable, varying only between 5%–10%.

The groundwater level in Jizzakh Province has also been elevated by anthropogenic activities, which have led to additional water inputs into the groundwater zone. Extensive leaching during spring, high-intensity irrigation during the vegetative season, leakage from irrigation canals and drainage water collectors, and waterlogging due to damaged or blocked drainage networks are the main reasons for raised groundwater levels.

Fig. 1 Dynamics of groundwater (GW) level changes in irrigated land in Jizzakh Province

This can be problematic for crop growth because shallow saline groundwater levels can lead to reduced water uptake (e.g., in cotton) and thus to crop losses (Hutmacher et al., 1996; SANIIRI, 2005). Studies conducted by the Central Asian Research Institute of Irrigation (SANIIRI) have shown that cotton productivity decreases by 15%–20% in slightly saline soils, 35%–40% in moderately

saline soils, and 70%-80% in highly saline soils. In addition to reduced yields (from 2.10 to 0.90-1.25 t/hm²), salinity has also been shown to influence the quality of raw cotton fibers (SANIIRI, 2005).

4.2 Dynamics of Groundwater Mineralization Fluctuations in the Province

Groundwater observations for 1995-2016 indicated that the average groundwater mineralization in most of the irrigated land of the province fluctuated between 1.1-5.0 g/L (75.3% of all irrigated land), with an average of less than 1.0 g/L in 13.1% of the irrigated land and a range of 5.1-10.0 g/L in 10.5% of the irrigated land (Table 6).

Table 6 Dynamics of the long-term (1995-2016) average GW mineralization in Jizzakh Province

Mineralization (g/L)	Irrigated land area ($\times 10^3$ hm ²)	Percentage (%)
10.0		
Total		

In 1995, the irrigated area with groundwater mineralization of 0.0-3.0 g/L was 73.65×10^3 hm², while in 2016 the corresponding figure was 152.76×10^3 hm², representing an increase of 79.11×10^3 hm² (25.30% of all irrigated land). The area with average mineralization (3.1-5.0 g/L) also increased from 86.11×10^3 hm² in 1995 to 133.33×10^3 hm² in 2016, representing an increase of 47.22×10^3 hm² (14.57%).

In contrast, in 1995 the irrigated area with high groundwater mineralization (5.0-10.0 g/L) was 118.93×10^3 hm², while in 2016 it was 13.75×10^3 hm², representing a considerable decrease of 105.18×10^3 hm² (36.58%). Similarly, the irrigated area with groundwater mineralization of 10.0 g/L and above was reduced from 10.32×10^3 hm² in 1995 to 0.70×10^3 hm² in 2016 (97.00% reduction, Fig. 2 [Figure 2: see original paper]). In most of the irrigated land area of the province (75.33%), groundwater mineralization was in the range of 1.1 to 3.0-5.0 g/L.

Within the period 1995-2016, the area with relatively low groundwater mineralization increased from 58.70×10^3 to 91.00×10^3 hm², indicating slight improvement in the state of irrigated land in the province. During the same period, the area of irrigated land with average groundwater mineralization (3.1-5.0 g/L) decreased from 44.90×10^3 to 16.50×10^3 hm² (20.80%). This indicates positive changes in the dynamics of groundwater mineralization in the irrigated areas of the province.

In 1995, the area with low groundwater mineralization (0.0–3.0 g/L) was 73.65×10^3 hm², while in 2016 it was 152.76×10^3 hm², representing an increase of 79.11×10^3 hm² (25.34%). The area with average groundwater mineralization of 3.1–5.0 g/L was 86.11×10^3 hm² in 1995 and 133.33×10^3 hm² in 2016, representing an increase of 47.22×10^3 hm² (14.57%). The irrigated area with high groundwater mineralization (5.0–10 g/L) decreased from 118.93×10^3 to 13.75×10^3 hm² during 1995–2016, and therefore GW quality was improved over 105.18×10^3 hm² (36.58%). The area of irrigated land with average groundwater mineralization of 10.0 g/L or more decreased from 10.32×10^3 hm² in 1995 to 0.70×10^3 hm² in 2016 (a decrease of 9.62×10^3 hm² or 97.00%). The area with high groundwater mineralization (5.0–10.0 g/L) was 118.93×10^3 hm² in 1995, but this decreased to 13.75×10^3 hm² in 2016, meaning groundwater quality was improved over an area of 105.18×10^3 hm² (63.00%).

Thus, based on these analyses it can be concluded that the quality of groundwater in Jizzakh Province has considerably improved in recent years. One of the main causes of the wedging of groundwater outflow in the province is the improvement of the drainage network in the Jizzakh irrigation area, which has diverted drainage water toward the Aydar-Arnasay and Tuzkan lakes in the Mirzachul Desert in the southern part of the province.

Fig. 2 Long-term (1995–2016) dynamics of groundwater (GW) mineralization of irrigated land in Jizzakh Province

4.3 Salinization Level of Irrigated Soils in the Province

As shown in Figure 1, the majority of the irrigated area in Jizzakh Province was characterized by a groundwater level of more than 2.0 m. In addition, the majority of the irrigated area in the province was characterized by low groundwater mineralization (Fig. 2). However, because the availability of low-mineralized deep groundwater is limited, the additional water needed for crop growth in the irrigated areas of the province is provided by water diversion from the Syrdarya and Zarafshan rivers. The level of mineralization in the water of both these rivers has increased (Table 2), with the result that extensive irrigation has led not only to rising groundwater levels but also increased groundwater mineralization.

An efficient drainage network would prevent salt accumulation, but the technical infrastructure used in the drainage system in Jizzakh Province is outdated and not always well maintained; therefore, salt removal is incomplete.

The results of this study indicated that most of the irrigated land in Jizzakh Province is slightly saline. On average, 51.3% of the irrigated area fell into this category (Fig. 3 [Figure 3: see original paper]). Another 17.7% showed no sign of salinization. However, 29.0% of the irrigated area was found to be moderately saline and 2.0% was highly saline. Overall, salinization in Jizzakh Province was

much lower than in neighboring Bukhara Province, where 27.6% of the irrigated area is classified as moderately saline and 7.8% as highly saline (Kulmatov et al., 2015).

Fig. 3 Temporal soil salinization dynamics of the irrigated area in Jizzakh Province from 1995 to 2016

All relevant parameters (irrigation water mineralization, groundwater level, and groundwater salinity) are subject to considerable temporal dynamics (Table 2; Figs. 1 and 2), which means their influence on soil salinity is also dynamic. Therefore, managing salinization of irrigated land is a complex process requiring flexible agrotechnical practices. In recent years, this paradigm has been increasingly incorporated into the agrarian sector, explaining the observed improvements in salinity conditions in Jizzakh Province.

Salinization of agricultural land negatively affects soil structure and biological activity of soil microorganisms. This can lead to reduced utilization of mineral fertilizers by plants and may cause various plant diseases (SCNP, 2013). Cotton productivity in Uzbekistan has decreased by 15%-20% in areas with low-salinity soils, 35%-40% in areas with moderate-salinity soils, and 70%-80% in areas with high-salinity soils (SANIIRI, 2005). Therefore, monitoring and assessment of salinization levels in irrigated soils is very important for agriculture. Based on the results obtained, programs to control the effects of salt leaching on the following year's crop can be initiated.

In Jizzakh Province in 1995, the area of irrigated land not affected by salinization was 50.57×10^3 hm² (17.50% of total irrigated land area), which increased to 67.64×10^3 hm² (22.50%; Fig. 3) in 2016. This indicates that in recent years, the irrigated area not affected by salinization increased by 17.07×10^3 hm² (5.00%). The area of low-salinity irrigated land in 1995 was 159.81×10^3 hm² (55.30%), which decreased to 158.67×10^3 hm² (52.80%) in 2016, representing a slight decrease of 1.15×10^3 hm² (0.50%). The area of moderate-salinity irrigated land in 1995 was 74.59×10^3 hm² (25.80%), which decreased to 68.62×10^3 hm² (22.80%) in 2016, representing a decrease of 5.97×10^3 hm² (3.00%). The high-salinity irrigated land area slightly increased from 4.04×10^3 hm² (1.40%) in 1995 to 5.62×10^3 hm² (2.00%) in 2016.

In general, it can be concluded that during the period 1995-2016, the salinization level of irrigated lands in the province has slightly increased. Based on the average soil salinization level, the province's irrigated areas could be classified as follows: no salinity (17.7%), low salinity (51.3%), moderate salinity (29.0%), and high salinity (2.0%) (Fig. 3). The observed slight increase of salinity in irrigated land in Jizzakh Province occurred due to the use of outdated irrigation technology and inefficient drainage systems. Because of relatively high amounts of mineralized salts in irrigated land and the proximity of groundwater, reclamation leaching of irrigated land is required each winter.

4.4 “Irrigation Water-Soil-Drainage Water” Salt Balance

Climate change and frequently recurring droughts in Central Asia in recent years have diminished river water flows, with the result that more water resources are required for irrigation (SCNP, 2013). If the amount of water available for irrigation does not meet the scientifically based needs of the crop, or if excessive amounts of water are used, salts accumulate in irrigated soils. The accumulation of salts associated with irrigated agriculture is mainly caused by two factors: the introduction of salts in irrigation water to the land, and the extremely high level of saline groundwater caused by excessive irrigation and poor drainage systems (Dukhovny, 1983; Kovda, 2008). It is difficult for excess water to flow out naturally from irrigated areas, which leads to salinization and swamping and reduces soil fertility (Kulmatov et al., 2015, 2018). Increased salinity in water and soil has had serious negative impacts on agriculture, food security, and human health worldwide. The direct impact of salinity on both the soil microbial community and crops can seriously affect crop production, yield, and cropping patterns (Yan et al., 2015; Rahaman et al., 2019).

It is therefore important to identify the total amount of salts entering irrigated land with irrigation water, the amount of salt remaining in irrigated areas, and the amount of salt leached out through the drainage system (reclamation leaching). An analysis of the dynamics of the “irrigation water-soil-drainage water” salt balance in the irrigated area of the province was conducted based on monitoring data for 2000–2016 (Table 7).

During the period 2000–2016, an average of $2,690.9 \times 10^3 \text{ m}^3$ of water was used for irrigation during the vegetation growth period, autumn crop irrigation, and winter salt leaching in the province. The highest volume of irrigation water ($3,368.8 \times 10^3 \text{ m}^3$) was used in 2005, from which $799.0 \times 10^3 \text{ m}^3$ of water was diverted to the drainage systems (Table 7).

The average annual amount of water used for irrigation ($2,690.9 \times 10^3 \text{ m}^3$) had a total salt content of $3,271.4 \times 10^3 \text{ t}$, with chlorides accounting for $974.9 \times 10^3 \text{ t}$. The annual average amount of water flowing out as drainage water was $826.4 \times 10^3 \text{ m}^3$, which contained a total salt content of $2,878.0 \times 10^3 \text{ t}$, with chlorides accounting for $887.5 \times 10^3 \text{ t}$. Thus, the total amount of salts that remained in the irrigated area was $393.4 \times 10^3 \text{ t}$, with chlorides accounting for $87.4 \times 10^3 \text{ t}$.

On average, for each hectare of irrigated land in the province, 1.21 t of salts entered the soil with irrigation water and 1.06 t of salts were leached out by drainage water. Thus, 12.0% of the salts and 8.9% of the chlorides that entered the soil with irrigation water remained in the irrigated areas of the province. On average, 0.15 t of salt remained on each hectare of irrigated land (Table 7). The lowest amount of irrigation water used was $2,351.8 \times 10^3 \text{ m}^3$ in 2001 (an extreme drought year in the country and province), $980.6 \times 10^3 \text{ m}^3$ of which was diverted to the drainage system. The remaining $1,371.2 \times 10^3 \text{ m}^3$ of water was used in the vegetation growth season. In autumn-winter, $88.9 \times 10^3 \text{ m}^3$ of river water is used each year for reclamation leaching in irrigated soils in the province.

This has resulted in desalinization and improvement in the reclamation status of irrigated land in the province.

The average salt concentration in water used in the province for irrigation purposes was 1.2 g/L in 2000–2016. The maximum amount of salts ($4,099.2 \times 10^3$ t) that entered the soils of the irrigated area occurred in 2000, of which $4,319.1 \times 10^3$ t was leached out by drainage water. Thus, the amount of salt entering the irrigated area in irrigation water was less than the amount of salt leached out from the irrigated area by drainage water. This phenomenon likely occurred because salt leaching in the irrigated areas in the autumn–winter period was achieved more effectively in 2000 than in other years.

In 2012, the total amount of salt entering the soil through irrigation water was $2,865.8 \times 10^3$ t, of which $2,209.1 \times 10^3$ t was leached out by drainage water. Thus, only 656.7×10^3 t of salts remained in the irrigated area. In 2000–2016, the average salt content in irrigation water was $3,271.4 \times 10^3$ t, of which $2,878.0 \times 10^3$ t was leached out from irrigated land. The total amount of residual salt in the irrigated area was 393.4×10^3 t. As a result, 1.3 t/hm² of salt has accumulated in irrigated land during the 17 years studied. Thus, 11.7% of the total incoming salts have remained in the irrigated area (Table 7).

Table 7 “Irrigation water–soil–drainage water” salt balance of Jizzakh Province during 2000–2016

Parameter	Irrigation water input	Drainage water output	Salt balance
	($\times 10^3$ m ³)	(g/L)	($\times 10^3$ t)
Average			

Note: WV, water volume; M, mineralization; C, chlorine content; TS, total salt content; TC, total chlorine content; TSB, total salt balance; TCB, total chlorine balance.

During the period 2000–2016, the average chlorine concentration in water used for irrigation in the province was 0.3 g/L. The lowest amount of chlorine that entered irrigated fields (823.2×10^3 t) was observed in 2014, with 886.3×10^3 t of chlorine leached out through drainage waters. This implies that the amount of chlorine in water used for irrigation was less than the amount flowing out in drainage water. The likely reason for this was that soil leaching was more effectively conducted during the autumn–winter season in 2014 than in other years.

In 2013, the total amount of chlorine entering soils in irrigation water was $1,364.3 \times 10^3$ t, of which 895.4×10^3 t was leached out through drainage water. As a result, 468.9×10^3 t of excess chlorine remained in irrigated areas.

In general, during 2000–2016, the average amount of chlorides in water used for irrigation was 974.9×10^3 t/a, of which 887.5×10^3 t/a was leached out through

drainage water. The amount of residual chlorides was 87.4×10^3 t/a, with 0.29 t/a of chlorides per hectare of irrigated land accumulating during the 17-year study period. On average, 10.1% of the total incoming chlorides accumulated in irrigated land (Table 7). Thus, analysis of the “irrigation water-soil-drainage water” salt balance indicated a low salinization level in the irrigated soils of the region.

Most salt-affected land has experienced salinization due to natural causes, with accumulation of salts over long periods in arid and semiarid zones (Rengasamy, 2002). More than 800×10^4 hm² of land throughout the world are salt-affected, including 4.5×10^4 hm² (20%) of the current 2.3×10^5 hm² of irrigated land (FAO, 2008). We found that the average value of groundwater salinization in most (75.3%) of the irrigated land of Jizzakh Province fluctuated between 1.1–5.0 g/L, with values less than 1.0 g/L in 13.1% of the irrigated land and in the range of 5.1–10.0 g/L in 10.5% of the irrigated land. High groundwater salinization levels can lead to increased salt content in the upper layers of soil through capillary action, which affects the root zone of crops, reducing their productivity. Salinization can be effectively controlled in irrigated land by reducing the groundwater level, increasing the efficiency of drainage networks that divert drainage water, and most importantly, using modern climate-smart water-saving irrigation techniques.

Low rainfall in dryland areas, high transpiration rates by vegetation, and high evaporation rates during summer cause accumulation of salts in root zone layers (Rengasamy, 2002). Temperatures in the ASB are likely to increase by 2°C–3°C by 2050 and 3°C–5°C by 2080, and are predicted to be particularly high in summer and autumn (Lioubimtseva, 2015). Under elevated soil salinity, repeated remediation leaching on irrigated land reduces surface soil salinity, but because existing drainage systems do not cover all irrigated land area and are inefficient, permanent salinization, decline in fertility, and reduced yields of agricultural crops can occur.

During the period 1995–2016, the salinization level of irrigated lands in Jizzakh Province increased slightly, and the area could be divided into the following classes: no salinity (17.7% of total area), low salinity (51.3%), moderate salinity (29.0%), and high salinity (2.0%). Remediation leaching of soils with elevated salinity requires use of additional water resources, which increases groundwater level and mineralization and also accelerates the salinization process in irrigated soils.

5 Conclusions

Under current conditions of low soil fertility, stable crop yields (e.g., 10%–15% for cotton and grain) in Jizzakh Province have been achieved mainly due to the use of large volumes of mineral fertilizers. However, permanent use of excess

amounts of mineral fertilizers (primarily nitrogen-containing) can also cause soil salinization.

The average value of groundwater salinization in most (75.3%) of the irrigated land fluctuated between 1.1–5.0 g/L, with values less than 1.0 g/L in 13.1% of the irrigated land and in the range of 5.1–10.0 g/L in 10.5% of the irrigated land. The condition of irrigated land in the province has not improved significantly. High groundwater salinization levels can lead to reduced crop productivity. The best way to effectively control salinization is to reduce the groundwater level, increase the efficiency of drainage networks, and most importantly, use modern climate-smart irrigation techniques.

The salt balance in the “irrigation water–soil–drainage water” system is extremely complex. It changes over time and requires monitoring and analysis of a large number of parameters. Repeated remediation leaching reduces surface soil salinity, but because existing drainage systems do not fully cover the irrigated land area and are ineffective, permanent salinization, decline in fertility, and reduced yields of agricultural crops are unavoidable. During the period 1995–2016, the salinization level of irrigated land in Jizzakh Province has slightly increased, and the area could be divided into the following classes: no salinity (17.7% of total area), low salinity (51.3%), moderate salinity (29.0%), and high salinity (2.0%).

Analysis of the long-term monitoring data for Jizzakh Province revealed the importance of sustainable use of available water resources, as well as the need to improve the reclamation status of irrigated land, maintain groundwater level and mineralization at optimal levels, keep irrigated soil salinization at optimal levels, and maintain the drainage system to ensure its effective operation. This would enable sustainable use of water and land resources under a changing climate. Detailed studies of the salt balance in irrigated areas, and the impact of climate change, increased fertilizer use, and repeated remediation leaching on groundwater level and mineralization, should be conducted in the future to better understand the processes leading to accelerated salinization, declines in fertility, and reduced yields of agricultural crops.

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