

Meta-Analysis of N₂O Emissions from Chinese Cropland Soils Under Biochar Application: Post-print

Authors: Luo Xiaoqi, Feng Hao, Liu Jingjing, Zhang Afeng

Date: 2017-11-10T00:00:00+00:00

Abstract

To clarify the effects of biochar application on N₂O emissions from Chinese cropland soils and identify the main controlling factors, published experimental data were collected and analyzed using meta-analysis to quantitatively evaluate the impacts of climate, soil properties, field management practices, biochar characteristics, and application rate on soil N₂O emissions under biochar amendment, with path analysis performed for each influencing factor. The results showed that when annual rainfall $\leq 600\text{mm}$, biochar significantly reduced soil N₂O emissions ($P < 0.05$), and this mitigation effect strengthened with increasing annual rainfall; when annual sunshine hours $> 1000\text{h}$, the N₂O emission reduction effect of biochar weakened with increasing sunshine hours. When soil pH ≤ 6.5 , the N₂O mitigation effect of biochar exhibited a trend of initially increasing then decreasing with rising soil pH; biochar application in loam soil produced a significant N₂O emission reduction ($P < 0.05$), whereas the effects in sandy and clay soils were not significant ($P > 0.05$). The N₂O emission reduction effect of biochar was superior in mulched soils compared to non-mulched soils; the mitigation effect weakened with increasing nitrogen fertilizer application rate but strengthened with increasing biochar specific surface area. When biochar C/N ratio ranged from 30–500, soil N₂O emissions under biochar application were significantly reduced ($P < 0.05$); when biochar application rate ranged from 20–160 t hm⁻², the N₂O emission reduction effect of biochar strengthened with increasing application rate. The effect of biochar on soil N₂O emission reduction displayed significant regional characteristics, with significant effects observed in South China, East China, Central China, and Northeast China ($P < 0.05$), but not in Northwest China ($P > 0.05$). Nitrogen fertilizer application rate, biochar application rate, mean annual temperature, and annual rainfall were the most critical factors influencing the emission reduction effectiveness of biochar, with interactions among these factors jointly affecting the N₂O mitigation effect of biochar. This study can provide a reference for the

promotion and application of biochar in agricultural regions of China and for cropland N₂O emission reduction.

Full Text

Abstract

To clarify the effects of biochar application on soil N₂O emissions and identify the main controlling factors in Chinese farmland, this study employed Meta-analysis to quantitatively analyze the impacts of climate, soil properties, field management practices, and biochar characteristics (including application rate) on soil N₂O emissions. Path analysis was subsequently used to determine the dominant influencing factors. The results demonstrated that biochar application significantly reduced soil N₂O emissions when annual precipitation was $\geq 600\text{mm}$ ($P < 0.05$), with the mitigation effect strengthening as precipitation increased. When annual sunshine hours exceeded 2000h ($P < 0.05$) but not in sandy or clay soils ($P > 0.05$). The mitigation effect was more pronounced in mulched soils than in non-mulched soils and diminished with increasing nitrogen application rate while strengthening with increasing biochar application rate ($P < 0.05$). At application rates of $20\text{--}160\text{t}\cdot\text{hm}^{-2}$, the mitigation effect intensified with increasing biochar application rate ($P < 0.05$) but no significant effect in Northwest China ($P > 0.05$). Nitrogen fertilizer input, biochar application amount, and soil N₂O emissions.

1.1 Data Sources

Literature published before December 2016 was systematically searched from Chinese and English databases including CNKI, VIP, Wanfang, Web of Science, and Google Scholar using keywords such as “biochar,” “black carbon,” “nitrous oxide,” “N₂O,” “charcoal,” and related terms. The screening criteria were: (1) studies conducted on Chinese farmland soils; (2) field, plot, or pot experiments; (3) at least one paired treatment of biochar application and control with identical field conditions; (4) clear documentation of experimental location, duration, and basic properties of soil and biochar; (5) provision of soil N₂O emission data or N₂O-N equivalents; (6) minimum of three replicates per treatment; and (7) selection of the longest-duration study when multiple publications reported identical experimental sites, years, and crop types. Based on these criteria, 41 valid publications were obtained, yielding 132 datasets.

1.2 Data Classification

Data were grouped according to factors influencing soil N₂O emissions, including experimental region, climatic factors (annual precipitation, mean temperature, sunshine hours), soil properties (pH and texture), field management practices (nitrogen application rate and mulching), and biochar characteristics (production temperature, specific surface area, C/N ratio) and application rate. Classification standards followed reference [15]. Using China’s regional classification method [16], study sites were divided into five regions: East China (Jiangsu, Zhe-

jiang, Jiangxi, Anhui), Central China (Hunan, Henan), South China (Guangdong), Northwest China (Xinjiang, Shaanxi, Ningxia), and Northeast China (Liaoning). Annual precipitation thresholds were set at 400 mm and 600 mm [17], and mean annual temperature at 10 °C [18]. Data categories are summarized in .

1.3 Meta-analysis

Since some studies reported N₂O emissions as CO₂ equivalents or N₂O-N, data conversion was necessary. For CO₂ equivalents, values were divided by 298 to obtain N₂O emissions [19]. For N₂O-N values, conversion used a molecular weight factor of 28/44 [15]. Standard deviation, a critical parameter in Meta-analysis representing study weights, was obtained either directly from publications or by digitizing graphical data using Origin 9.0. When standard deviations were not reported, MetaWin 2.1 software's resampling function was used to generate unweighted variance estimates [21-22].

Meta-analysis was performed using MetaWin 2.1, requiring mean N₂O emissions, standard deviations, and sample sizes for both biochar and control treatments. Effect size was quantified using a random-effects model to calculate the response ratio (lnR) [23]:

$$\ln R = \ln \left(\frac{X_e}{X_c} \right)$$

where R is the response ratio (N₂O emissions with biochar, X_e, divided by emissions without biochar, X_c). To express the direction and magnitude of biochar effects, the percentage change in N₂O emissions was calculated as:

$$\text{Change rate} = (R - 1) \times 100\%$$

The 95% confidence intervals were calculated following reference [24]. If the interval included 0, biochar had no significant effect (P>0.05). If entirely above 0, biochar significantly promoted emissions (P<0.05). If entirely below 0, biochar significantly reduced emissions (P<0.05) [25].

1.4 Path Analysis

Path analysis is a multivariate statistical method that reveals relationships among variables by partitioning direct, indirect, and total effects. Direct path coefficients represent a factor's immediate influence on soil N₂O emissions, indirect coefficients show effects mediated through other variables, and total coefficients indicate overall impact [26]. Computational procedures followed reference [27].

1.5 Data Processing

Microsoft Excel 2010 was used for database construction, Origin 9.0 for figure preparation and digitization, MetaWin 2.1 for Meta-analysis, and SPSS 19.0 for path analysis.

2.1 Effects of Climatic Factors on Soil N₂O Emissions

Analysis of 41 publications revealed that biochar's mitigation effect on soil N₂O emissions was closely related to climatic conditions [Figure 1: see original paper].

2.1.1 Effects of Annual Precipitation

Meta-analysis showed that annual precipitation attenuated soil N₂O emissions in Chinese farmland. When precipitation was <600 mm, biochar had no significant mitigation effect ($P>0.05$). However, at \$600 mm, biochar significantly reduced emissions ($P<0.05$). The reduction rates were 10.7% (95% CI: 1.1% to -22.5%), 19.2% (95% CI: 2.4% to -40.8%), and 25.3% (95% CI: -6.6% to -43.4%) for precipitation levels of <400 mm, 400-600 mm, and \$600mm, respectively, indicating that mitigation strengthened with increasing precipitation. Higher precipitation reductase [28]. Additionally, N₂O diffusion from deep soil to the atmosphere decreases with higher water content, while nitrifier activity declines, limiting substrates for denitrification. Biochar adsorbs ammonium, reducing energy substrates for nitrifiers, and its aromatic carbon structure adsorbs substrates for denitrifiers, thereby inhibiting both microbial groups and enhancing N₂O mitigation [29].

2.1.2 Effects of Annual Mean Temperature and Sunshine Hours

The 10 °C threshold represents the temperature at which warm-season plants begin accumulating organic matter [30]. At temperatures <10 °C and \$10°C, biochar significantly reduced N₂O emissions by 27.3% and 17.9%, respectively, showing that mitigation effectiveness decreased with rising temperature. For sunshine hours of 1,000-2,000 h, the reduction rate was 21.3%, while at \$2,000hitwas16.7{2}\$O mitigation.

Overall, biochar's mitigation effect diminished with increasing temperature and sunshine hours. Both factors influence soil organic matter decomposition, microbial enzyme activity, and crop growth [31]. Higher temperatures enhance microbial and enzyme activities, strengthening substrate uptake by nitrifiers and denitrifiers and promoting organic matter mineralization [32]. Sunshine duration affects both soil temperature and crop growth, influencing N₂O emissions. Extended sunshine enhances root water and nutrient uptake, root oxygen consumption, and organic exudation, promoting denitrification [33]. Roots also absorb dissolved N₂O that cannot diffuse to the atmosphere, releasing it via transpiration [34]. Although biochar reduces N₂O emissions, its adsorption capacity

is limited, so mitigation effectiveness declines as temperature and sunshine increase.

2.2 Effects of Soil Properties

2.2.1 Effects of Soil pH

Soil pH affects nitrifier and denitrifier activities, altering rates and end products of nitrification and denitrification [35]. At pH<6.5, biochar reduced N₂O emissions by 31.5%, but not significantly (P>0.05). At pH 6.5-7.5 and \$ 7.5, reductions were significant (P < 0.05) at 5.23 [2] Omissions [36]. In acidic soils, microbial metabolism is reduced, containing functional groups show reduced heavy metal adsorption, impairing microbial habitat, and weakened adsorption [31,38].

2.2.2 Effects of Soil Texture

Biochar significantly reduced N₂O emissions in loam soils (P<0.05) but not in sandy or clay soils (P>0.05). Reduction rates were 2.8%, 22.6%, and 13.7% for sandy, loam, and clay soils, respectively, with loam showing the best mitigation. Soil texture influences nitrification, denitrification, N₂O diffusion, and organic matter decomposition [39-40]. Gas diffusion coefficients vary with texture, and loam's balanced pore structure facilitates N₂O emission pathways, resulting in higher baseline emissions than sandy or clay soils [41].

2.3 Effects of Field Management Practices

2.3.1 Effects of Nitrogen Fertilizer Application Rate

Nitrogen application increases soil nitrogen content and substrates for nitrification (NH₄⁺ and NO₃⁻) while stimulating root growth and exudation, affecting N₂O emissions [42]. At application rates \$ 60 kg · hm⁻², biochar reduced emissions but not significantly (P > 0.05). The mitigation effect decreased with increasing N:P ratio in denitrification products [30]. Although biochar improves soil microenvironment and reduces nitrogen-related microbial abundance [44], its adsorption capacity is limited, causing mitigation effectiveness to decline at high nitrogen rates.

2.3.2 Effects of Plastic Film Mulching

Mulching alters soil temperature and moisture, affecting N₂O production, transport, and emissions. Biochar significantly reduced N₂O emissions under both mulched and non-mulched conditions (P<0.05), with a 25.3% reduction in mulched soils—6.2% greater than in non-mulched soils. Plastic film physically blocks water evaporation, increasing soil moisture [45] and promoting organic matter mineralization and denitrification. Film also absorbs solar radiation, reducing heat exchange and increasing soil temperature, which enhances microbial activity and organic matter decomposition [46]. Mulching concentrates

N₂O in a confined humid space, increasing its concentration and biochar adsorption efficiency, thereby improving mitigation effectiveness.

2.4 Effects of Biochar Properties and Application Rate

2.4.1 Effects of Biochar Production Temperature, Specific Surface Area, and C/N Ratio

Biochar production temperature, specific surface area, and C/N ratio all influence N₂O emissions. Production temperature significantly reduced emissions ($P < 0.05$), with reduction rates of 32.6%, 43.9%, 35.6%, and 51.3% for temperatures of < 400 °C, 400–500 °C, 500–600 °C, and ≥ 600 °C, respectively. This aligns with Cayuela et al. [47] because surface functional groups change with temperature; higher temperatures increase aromaticity, reduce polarity, and enhance stability [48]. Low-temperature biochar has uniform micropores, while elevated temperatures cause irregular distribution and pore wall collapse, increasing surface roughness [49], consistent with observed emission patterns.

Specific surface area showed no significant overall effect ($P > 0.05$), but reduction rates were 15.5% and 35.1% for areas < 100 and $\geq 100 \text{ m}^2 \text{ g}^{-1}$, respectively, indicating that mitigation improves with increasing surface area. Larger surface area enhances adsorption capacity [50].

When biochar C/N ratio was 10–30, the effect was not significant ($P > 0.05$), but at 30–500 it became significant ($P < 0.05$). Mitigation increased then stabilized with higher C/N ratios because elevated C/N intensifies microbial activity while limiting nitrogen supply for nitrification and denitrification, promoting microbial immobilization of mineral nitrogen and interspecies competition, thereby reducing N₂O emissions [51].

2.4.2 Effects of Biochar Application Rate

Higher biochar rates increase adsorption area for ammonium and alter electron acceptors and redox potential [52], affecting N₂O emissions. At application rates of $20\text{--}160 \text{ t} \cdot \text{hm}^{-2}$, biochar significantly reduced emissions ($P < 0.05$), with mitigation strengthening as application rate increased. Rates below $20 \text{ t} \cdot \text{hm}^{-2}$ showed no significant effect ($P > 0.05$). Increased biochar application enhances adsorption of nitrification and denitrification substrates, reducing microbial N₂O production [53]. New porosity created by biochar improves anaerobic microbial community function and diversity, reducing denitrifier populations and soil denitrification potential [54], consistent with findings by Li et al. [55] and Jia et al. [56].

2.5 Regional Characteristics of Biochar Effects on Soil N₂O Emissions

Biochar effects varied significantly across China [Figure 5: see original paper]. In South, East, Central, and Northeast China, biochar significantly reduced

N_2O emissions ($P < 0.05$) by 31.4%, 27.2%, 26.7%, and 21.5%, respectively. In Northwest China, emissions decreased by 17.2%, but not significantly ($P > 0.5$). Biochar showed the best mitigation in South China and the worst in Northwest China, correlating with regional climate, soil properties, and management practices. Annual precipitation decreases from southeast to northwest, with South China receiving the most rainfall [57]; since mitigation strengthens with precipitation, South China showed the greatest effect. Spatial patterns of sunshine hours (more in north and west) and temperature variations also influenced regional effectiveness. Soil property variability and different cropping systems further contributed to these regional patterns [59-60], as did management practices affecting soil hydrothermal conditions and microbial communities [61].

2.6 Path Analysis of Biochar Effects on Soil N_2O Emissions

Since biochar mitigation is influenced by multiple interacting factors, path analysis was conducted using quantitative parameters: annual precipitation (X_1), mean temperature (X_2), sunshine hours (X_3), soil pH (X_4), texture (X_5), nitrogen rate (X_6), biochar production temperature (X_7), specific surface area (X_8), C/N ratio (X_9), application rate (X_{10}), and N_2O emission reduction rate (I) .

Direct path coefficients ranked as: $X_6 > X_{10} > X_2 > X_1 > X_3 > X_9 > X_4 > X_5 > X_8 > X_7$, indicating that nitrogen fertilizer input, biochar application rate, and mean temperature were the three dominant factors. Annual precipitation and sunshine hours had similar direct effects, while biochar production temperature and specific surface area had smaller direct impacts. All factors showed positive direct effects on mitigation.

Indirect effects revealed that temperature and precipitation most strongly influenced mitigation indirectly through soil pH, while sunshine hours acted primarily through precipitation. Among soil properties, pH and texture interacted strongly, with each influencing mitigation primarily through the other. Nitrogen rate affected mitigation mainly through soil pH. Among biochar properties, production temperature acted primarily through specific surface area, while specific surface area, C/N ratio, and application rate all influenced mitigation mainly through soil pH. Total path coefficients showed that nitrogen fertilizer input, biochar application rate, mean temperature, and precipitation were the most important factors overall, with pH's indirect contributions being substantial. Therefore, optimizing nitrogen and biochar application while managing soil pH is crucial for effective N_2O mitigation in Chinese farmland.

Conclusions

- 1) Biochar application reduces soil N_2O emissions in Chinese farmland, but the magnitude depends on external factors including precipitation, temperature, sunshine hours, soil pH and texture, nitrogen application rate, and plastic film mulching.

- 2) Biochar properties including production temperature, specific surface area, C/N ratio, and application rate also affect emissions, with C/N ratio and application rate being particularly influential.
- 3) Biochar significantly reduced N₂O emissions in South, East, Central, and Northeast China, but not in Northwest China.

Therefore, achieving effective N₂O mitigation through biochar application requires integrated consideration of climate, soil properties, field management, and biochar characteristics.

References

- [1] UNEP. The emissions gap report[R]. Nairobi: United Nations Environment Programme, 2013
- [2] Zhao Z C, Han X, Shi Y F, et al. Effect of nitrification and urease inhibitor on carbon sequestration and greenhouse gas emissions in winter wheat and summer maize rotation system in North China[J]. Transactions of the Chinese Society of Agricultural Engineering, 2016, 32(6): 254-262
- [3] Lan Y, Meng J, Yang X, et al. Effects of different straw incorporation ways on N₂O emission and soil physicochemical properties of brown soil[J]. Chinese Journal of Ecology, 2015, 34(3): 790-796
- [4] Li B, Fan C H, Zhang H, et al. Combined effects of nitrogen fertilization and biochar on the net global warming potential, greenhouse gas intensity and net ecosystem economic budget in intensive vegetable agriculture in southeastern China[J]. Atmospheric Environment, 2015, 100: 10-19
- [5] Wang X, Yin D X, Zhang F, et al. Analysis of effect mechanism and risk of biochar on soil fertility and environmental quality[J]. Transactions of the CSEA, 2015, 31(4): 248-257
- [6] Karhu K, Mattila T, Bergström I, et al. Biochar addition to agricultural soil increased CH₄ uptake and water holding capacity-results from a short-term pilot field study[J]. Agriculture, Ecosystems & Environment, 2011, 140(1/2): 309-313
- [7] Nieder R, Benbi D K, Scherer H W. Fixation and defixation of ammonium in soils: A review[J]. Biology and Fertility of Soils, 2011, 47(1): 1-14
- [8] Zavalloni C, Alberti G, Biasiol S, et al. Microbial mineralization of biochar and wheat straw mixture in soil: A short-term study[J]. Applied Soil Ecology, 2011, 50: 45-51
- [9] He F F, Rong X M, Liang Y S, et al. Effects of biochar on soil physicochemical properties and N₂O, CO₂ emissions from vegetable-planting red soil[J]. Journal of Agro-Environment Science, 2013, 32(9): 1893-1900
- [10] Li B, Fan C H, Xiong Z Q, et al. The combined effects of nitrification inhibitor and biochar incorporation on yield-scaled N₂O emissions from an intensively managed vegetable field in southeastern China[J]. Biogeosciences, 2015, 12(6):
- [11] Liu X Y, Ye Y X, Liu Y M, et al. Sustainable biochar effects for low carbon crop production: A 5-crop season field experiment on a low fertility soil from

- Central China[J]. *Agricultural Systems*, 2014, 129: 22-29
- [12] Wang J, Shi Y, Li Z Y, et al. Effects of biochar application on N₂O emission in degraded vegetable soil and in remediation process of the soil[J]. *Acta Pedologica Sinica*, 2016, 53(3):
- [13] Hedges L V, Gurevitch J, Curtis P S. The meta-analysis of response ratios in experimental ecology[J]. *Ecology*, 1999, 80(4): 1150-1156
- [14] Challinor A J, Watson J, Lobell D B, et al. A meta-analysis of crop yield under climate change and adaptation[J]. *Nature Climate Change*, 2014, 4(4): 287-291
- [15] Zhang R, Zhao X, Pu C, et al. Meta-analysis on effects of residue retention on soil N₂O emissions and influence factors in China[J]. *Transactions of the Chinese Society of Agricultural Engineering*, 2015, 31(22): 1-6
- [16] Tian K, Zhao Y C, Xing Z, et al. A meta-analysis of long-term experiment data for characterizing the topsoil organic carbon changes under different conservation tillage in cropland of China[J]. *Acta Pedologica Sinica*, 2013, 50(3): 433-440
- [17] Zhang J T, Li Z. A study on demarcation indexes between subhumid and semiarid sectors in China[J]. *Progress in Geography*, 1999, 18(3): 230-237
- [18] Li F, Wang C, Zhao J, et al. The spatialization of multi-year average accumulated temperature in China[J]. *Journal of Natural Resources*, 2010, 25(5): 778-784
- [19] Stocker T F, Qin D, Plattner G K, et al. IPCC: Climate change 2013: The physical science basis[C]//Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change. Cambridge: Cambridge University Press, 2013: 710-719
- [20] Zhao A Q, Wei X J, Zhu M. Meta analysis on impact of plastic film on potato yield in China[J]. *Transactions of the CSAE*, 2015, 31(24): 1-7
- [21] Bax L, Yu L M, Ikeda N, et al. A systematic comparison of software dedicated to meta-analysis of causal studies[J]. *BMC Medical Research Methodology*, 2007, 7: 40
- [22] Yuan J L, Liang X Q, Li L, et al. Response of rice yield and nitrogen uptake to enhanced efficiency nitrogen fertilizer in China: A meta-analysis[J]. *Scientia Agricultura Sinica*, 2014, 47(17): 3414-3423
- [23] Zheng F Y, Peng S L. Comparison of two effect sizes of meta-analysis commonly used in ecology[J]. *Ecologic Science*, 2005, 24(3): 250-253
- [24] Zheng K, He J, Li H W, et al. Meta-analysis on maize and wheat yield under subsoiling in northern China[J]. *Transactions of the CSEA*, 2015, 31(22): 7-15
- [25] Morgan P B, Ainsworth E A, Long S P. How does elevated ozone impact soybean? A meta-analysis of photosynthesis, growth and yield[J]. *Plant, Cell & Environment*, 2003, 26(8):
- [26] Li X B, Wang Y K, Zhang P. Dynamic changes of pear jujube stem diameter and path analysis of MDS influencing factors with full irrigation[J]. *Transactions of the CSAE*, 2011, 27(4):
- [27] Sun S K, Wang Y B, Wu P T, et al. Spatial variability and attribution analysis of water footprint of wheat in China[J]. *Transactions of the Chinese*

Society of Agricultural Engineering, 2015, 31(13): 142-148

[28] Koponen H T, Martikainen P J. Soil water content and freezing temperature affect freeze-thaw related N₂O production in organic soil[J]. *Nutrient Cycling in Agroecosystems*, 2004, 69(3): 213-219

[29] Gundale M J, DeLuca T H. Temperature and source material influence ecological attributes of ponderosa pine and Douglas-fir charcoal[J]. *Forest Ecology and Management*, 2006, 231(1/3): 86-93

[30] Liu S H, Yan D H, Weng B S, et al. Spatiotemporal evolution of effective accumulated temperature $\geq 10^{\circ}\text{C}$ in China in recent 50 years[J]. *Arid Zone Research*, 2013, 30(4) :

689-696 [31] Zhu Y G, Wang X H, Yang X R, et al. Key microbial processes in nitrous oxide emission of agriculture

792-800 [32] Yang Y, Huang M, Liu H S, et al. The interrelation between temperature sensitivity and adaptability.

1811-1820 [33] Li B B, Zeng K, Li R, et al. A review on soil N₂O emission as influenced by crop growth[J]. *Chinese*

1003-1010 [34] Zhang Y M, Hu C S, Zhang J B, et al. Research advances on source/sink intensities and greenhouse

and N₂O in agricultural soils[J]. *Chinese Journal of Eco-Agriculture*, 2011, 19(4): 966-975

[35] Van Den Heuvel R N, Bakker S E, Jetten M S M, et al. Decreased N₂O reduction by low soil pH causes high N₂O emissions in a riparian ecosystem[J]. *Geobiology*, 2011, 9(3): 191-199

[36] Nägele W, Conrad R. Influence of soil pH on the nitrate-reducing microbial populations and their potential to reduce nitrate to NO and N₂O[J]. *FEMS Microbiology Letters*, 1990, 74(1): 49-57

[37] Luo Y, Durenkamp M, De Nobili M, et al. Short term soil priming effects and the mineralisation of biochar following its incorporation to soils of different pH[J]. *Soil Biology and Biochemistry*, 2011, 43(11): 2304-2314

[38] Yang W H, Weber K A, Silver W L. Nitrogen loss from soil through anaerobic ammonium oxidation coupled to iron reduction[J]. *Nature Geoscience*, 2012, 5(8): 538-541

[39] Xu H, Xing G X, Cai Z C, et al. Effect of soil water regime and soil texture on N₂O emission from rice paddy field[J]. *Acta Pedologica Sinica*, 2000, 37(4): 499-505

[40] Zhao M M, Zhang W Z, Pei Y, et al. Research advances on N₂O emission in agricultural soil[J]. *Crops*, 2013, (4): 25-31

[41] Rochette P, Angers D A, Chantigny M H, et al. N₂O fluxes in soils of contrasting textures fertilized with liquid and solid dairy cattle manures[J]. *Canadian Journal of Soil Science*, 2008, 88(2): 175-187

[42] Hansen S, Mæhlum J E, Bakken L R. N₂O and CH₄ fluxes in soil influenced by fertilization and tractor traffic[J]. *Soil Biology and Biochemistry*, 1993, 25(5): 621-630

[43] He F F, Jiang R F, Chen Q, et al. Nitrous oxide emissions from an intensively managed greenhouse vegetable cropping system in northern China[J]. *Environmental Pollution*, 2009, 157(5): 1666-1672

[44] Gu M Y, Liu H L, Li Z Q, et al. Impact of biochar application on soil nutrients and microbial diversities in continuous cultivated cotton fields in Xinjiang[J]. *Scientia Agricultura Sinica*, 2014, 47(20): 4128-4138

[45] Su W, Qu Y, Feng B L, et al. Photosynthesis characteristics and yield of broomcorn millet under film mulching on ridge-furrow for harvesting rainwater

- model in semi-arid region of northern Shaanxi[J]. Transactions of the CSAE, 2014, 30(13): 137-145
- [46] Li L L, Li F L, Liu Q Y, et al. Effect of plastic film mulching on the distribution and translocation of nitrogen in soil-lettuce system[J]. Acta Ecologica Sinica, 2011, 31(13): 3811-3819
- [47] Cayuela M L, Van Zwieten L, Singh B P, et al. Biochar' s role in mitigating soil nitrous oxide emissions: A review and meta-analysis[J]. Agriculture, Ecosystems & Environment, 2014, 191: 5-16
- [48] Lu H N, Hu X Y, Liu H W, et al. Influence of pyrolysis conditions on stability of biochar[J]. Environmental Science & Technology, 2013, 36(8): 11-14
- [49] Zhang J Y, Pu L J, Li G. Preparation of biochar adsorbent from straw and its adsorption capability[J]. Transactions of the CSAE, 2011, 27(S2): 104-109
- [50] Wang Z Y, Liu G C, Xing M, et al. Adsorption of aquatic Cd() varies with biochars derived at different pyrolysis temperature[J]. Environmental Science, 2014, 35(12): 4735-4744
- [51] Wang X Q, Li W S, Wang X D, et al. Greenhouse gases emissions and varies of carbon and nitrogen of manure from different lactating period cows[J]. Transactions of the Chinese Society for Agricultural Machinery, 2016, 47(3): 179-185
- [52] Singh B P, Hatton B J, Singh B, et al. Influence of biochars on nitrous oxide emission and nitrogen leaching from two contrasting soils[J]. Journal of Environmental Quality, 2010, 39(4): 1224-1235
- [53] Xu N N, Lin D S, Xu Y M, et al. Adsorption of aquatic Cd²⁺ by biochar obtained from corn stover[J]. Journal of Agro-Environment Science, 2014, 33(5): 958-964
- [54] Yanai Y, Toyota K, Okazaki M. Effects of charcoal addition on N₂O emissions from soil resulting from rewetting air-dried soil in short-term laboratory experiments[J]. Soil Science and Plant Nutrition, 2007, 53(2): 181-188
- [55] Li S, Li H L, Fang X B, et al. Biochar input to reduce trace greenhouse gas emission in paddy field[J]. Transactions of the CSAE, 2014, 30(21): 234-240
- [56] Jia J X, Xiong Z Q. Impact of application of maize stalk-derived biochar on soil properties of and N₂O, CO₂ and CH₄ emissions from vegetable fields[J]. Journal of Ecology and Rural Environment, 2016, 32(2): 283-288
- [57] Liu H. Statistical analysis on the precipitation and temperature over China in nearly 50 years and preliminary study on the impact of ocean[D]. Qingdao: Ocean University of China, 2009
- [58] Li H Q, Fu Z T, Wen X Y, et al. Characteristic analysis of sunshine duration change in China during the last 50 years[J]. Climatic and Environmental Research, 2013, 18(2): 203-209
- [59] Lei Z D, Yang S X, Xu Z R, et al. Preliminary investigation of the spatial variability of soil properties[J]. Journal of Hydraulic Engineering, 1985, (9): 10-21
- [60] Chen S T, Huang Y, Zheng X H, et al. Nitrous oxide emission from cropland and its driving factors under different crop rotations[J]. Scientia Agricultura Sinica, 2005, 38(10):

[61] Cheng C, Zeng Y J, Yang X X, et al. Effect of different tillage methods on net global warming potential and greenhouse gas intensity in double rice-cropping systems[J]. Acta Scientiae Circumstantiae, 2015, 35(6): 1887-1895

Note: Figure translations are in progress. See original paper for figures.

Source: ChinaXiv –Machine translation. Verify with original.